

Spatial models cannot compensate for sampling designs unstructured relative to isotopic variation: implications for baseline isoscape development in the Mediterranean Sea

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ABSTRACT

Isoscapes – spatial models of stable isotope ratios – are powerful tools for understanding biogeochemical cycles, food web structure, and ecological connectivity. In marine systems, however, broad-scale isoscapes are difficult to constrain because reference datasets are typically heterogeneous and opportunistic. Bayesian hierarchical models implemented through Integrated Nested Laplace Approximation (INLA) provide an efficient framework for modeling spatial patterns while accounting for non-spatial variance. We tested whether broad-scale geographic baseline isoscapes could be constrained for the Mediterranean Sea using INLA models applied to an opportunistically compiled dataset from this semi-enclosed basin, where strong environmental gradients are expected but spatially explicit baselines remain lacking. We assembled the most comprehensive dataset of $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ values to date for low-trophic-level, primarily epipelagic food web compartments ($> 6,000$ samples from > 600 sites). Although discrete regions of relatively high and low δ values were identified, particularly for $\delta^{15}\text{N}$, semivariograms and model outputs indicated only short-range spatial autocorrelation (ca. 40 km for $\delta^{13}\text{C}$; 65 km for $\delta^{15}\text{N}$) within this large dataset. Environmental covariates improved model fit for $\delta^{15}\text{N}$ values but did not enhance predictive performance or alter spatial predictions. Our results demonstrate that even advanced spatial models may not compensate where sampling design is structurally mismatched relative to unknown but structured spatial and non-spatial sources of isotopic variation. Coordinated sampling designed to capture spatial variability while accounting for major sources of non-spatial variance is required to develop robust marine isoscapes supporting ecological and management applications in the Mediterranean Sea and similar regions.

1. Introduction

Stable isotope ratios of elements often vary systematically across environments due to biogeochemical and physical processes (Bowen, 2010). These spatial patterns can be expressed as isoscapes – mechanistic or statistical models that explicitly predict isotopic variation across space (Bowen, 2010; West et al., 2010). Mechanistic models predict spatial patterns from underlying physical and biogeochemical processes, whereas statistical models infer spatial patterns directly from observed data. Isoscapes provide valuable insight into biogeochemical processes (Schmittner et al., 2013; Schmittner and Somes, 2016; Buchanan et al., 2019a, 2022) and have broad applications in spatial ecology (Vander Zanden et al., 2015; Trueman et al., 2017; St John Glew et al., 2019a; Trueman et al., 2019), trophic ecology (Navarro et al., 2013; Jennings and van der Molen, 2015; Bird et al., 2018; Pethybridge et al., 2018a), and forensic traceability (Chesson et al., 2010; Slessor and Trueman, 2021; Cusa et al., 2022; Martino et al., 2022).

In marine environments, isoscape development has traditionally lagged behind terrestrial systems due to the logistical challenges of obtaining reference samples at appropriate spatial and temporal scales (Magozzi et al., 2017; Trueman and St John Glew, 2019). Mechanistic models can partly overcome these limitations by predicting isotopic variation from first principles (Magozzi et al., 2017; Somes et al., 2010), but sample-based statistical isoscapes remain essential in systems where spatial isotope patterns emerge from multiple interacting processes that are difficult to resolve mechanistically.

Statistical isoscapes of carbon, nitrogen, and sulfur have been developed for several oceanic regions (Trueman et al., 2017; McMahon et al., 2013; MacKenzie et al., 2014; St John Glew et al., 2019b; Espinasse et al., 2020; St John Glew et al., 2021; Espinasse et al., 2022), offering opportunities for research and management at regional to basin scales. These isoscapes are typically generated by spatially interpolating measured isotope values based on geographic distance (Vander Zanden et al., 2015; Trueman et al., 2017), relationships with environmental

covariates (Jennings and Warr, 2003; Radabaugh et al., 2013), or both (MacKenzie et al., 2014). Ideally, reference samples are obtained through targeted, gridded sampling (e.g., Trueman et al., 2017) designed to capture spatially structured variation. In practice, however, most broad-scale marine isoscapes rely on opportunistically collected data, particularly for offshore or logistically challenging regions (McMahon et al., 2013; Espinasse et al., 2020; St John Glew et al., 2021; Espinasse et al., 2022). Such datasets are often unevenly distributed in space and heterogeneous with respect to time periods, depths, taxa, and analytical methods. This heterogeneity introduces non-spatial sources of isotopic variability – arising from temporal (seasonal to decadal scales; e.g., Espinasse et al., 2022; Schloesser et al., 2009; Bănarău et al., 2014; Liénart et al., 2022; Giraldo et al., 2024; Liénart et al., 2024), vertical (Saino and Hattori, 1980; Polunin et al., 2001; Koppelman et al., 2009; Fanelli et al., 2013; Pavlidou et al., 2020; Chaikalis et al., 2021), and biological (Vander Zanden and Rasmussen, 2001; McCutchan et al., 2003; Pecquerie et al., 2010; Canseco et al., 2022) processes – that can confound spatial patterns. As a result, when sampling is not structured relative to these sources of variation, underlying geographic signals may be obscured or difficult to resolve (St John Glew et al., 2019b).

The increasing availability of compiled isotope datasets (e.g., Arnoldi et al., 2023; Berto et al., 2025; IsoBank, 2025; IsoArcH, 2025) creates new opportunities to develop isoscapes. Addressing this challenge requires statistical approaches capable of partitioning variance into spatial and non-spatial components, allowing genuine geographic structure to be distinguished from variation arising from opportunistic data compilation (e.g., time periods, depths, taxa, sizes, trophic levels, methods). Bayesian hierarchical models implemented through Integrated Nested Laplace Approximation (INLA) provide an efficient and flexible framework for spatial modeling while accounting for systematic non-spatial variance, provided that spatial coverage is sufficient and that key sources of non-spatial variation can be constrained (St John Glew et al., 2019b, 2021; Espinasse et al., 2022). INLA-based methods therefore offer a means to test whether spatial variance can be reliably extracted from opportunistic datasets, and under what conditions spatial and non-spatial sources of variance can be distinguished. This is a question of broad relevance for large-scale marine isoscape

[†] Deceased.

development.

1.1. The Mediterranean Sea test case

The Mediterranean Sea is an ideal candidate for isoscape development. The region holds major ecological, economic, and geopolitical significance but faces intense anthropogenic pressure from coastal development, pollution, habitat degradation, and overfishing (Turley, 1999; Coll et al., 2010; Lotze et al., 2006; Halpern et al., 2008), with many fish stocks already overexploited (Hilborn et al., 2020). Rapid warming (Pisano et al., 2020; Kubin et al., 2023; Skliris et al., 2024; EEA Sea Surface Temperature, 2025) and pronounced tropicalization in the eastern basin (Vergés et al., 2014; Azzurro et al., 2019; Santana-Garcon et al., 2023) are reshaping community structure and ecosystem functioning. Lessepsian migrations through the Suez Canal continue to introduce non-indigenous species, further altering Mediterranean food webs (Raitos et al., 2010; Fanelli et al., 2015; Tiralongo et al., 2022). These ecological changes occur within a fragmented governance framework that complicates coordinated management and conservation (Mackelworth et al., 2019).

Stable isotope analyses have long been applied to ecological and biogeochemical studies in the Mediterranean, yet most applications remain geographically restricted and locally focused. Developing basin-scale isoscapes would provide a unifying, spatially explicit framework to integrate these diverse efforts, enabling isotopic data to be interpreted consistently across broader geographic contexts. In turn, this would enhance the capacity of isotope data to inform ecosystem-based management and spatial planning under accelerating environmental change. Table 1 summarizes key Mediterranean applications of stable isotopes that would directly benefit from spatialization through isoscapes.

The Mediterranean Sea consists of two deep basins with restricted circulation and deep-water exchange. It exhibits pronounced biogeochemical and climatic gradients: temperature and salinity increase eastward, while nutrient availability and productivity decrease (Coll et al., 2010; Salgado-Hernanz et al., 2019; Reale et al., 2022; Salgado-Hernanz et al., 2022; Álvarez et al., 2023). North-to-south gradients within both basins further reflect decreasing nutrient concentrations and productivity alongside increasing temperature and salinity. These environmental contrasts are expected to generate systematic isotopic variation in pelagic surface primary production, producing predictable $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ gradients in macronutrients and low-trophic-level food web compartments. For carbon, isotopic variability arises primarily from fractionation during photosynthesis (Rau et al., 1989, 1996, 1997), which depends on the carbon demand-to-supply ratio in phytoplankton and is therefore influenced by CO_2 availability, phytoplankton growth rate, and temperature (Rau et al., 1996, 1997; Laws et al., 1995, 1997). Consequently, longitudinal gradients in CO_2 concentration, nutrient availability, productivity, and temperature are expected to produce broad west-to-east differences in phytoplankton $\delta^{13}\text{C}$ values, as demonstrated in other marine regions (St John Glew et al., 2019b, 2021). For nitrogen, spatial patterns in $\delta^{15}\text{N}$ values reflect variation in nitrogen sources and utilization (Ryabenko, 2013): in oligotrophic eastern regions, enhanced N_2 fixation and atmospheric deposition are expected to introduce isotopically lighter nitrogen (Pantoja et al., 2002; Wald et al., 2025; Mara et al., 2009; Emeis et al., 2010; Krom et al., 2010), whereas in the more productive western regions, greater anthropogenic and terrestrial nitrogen inputs contribute ^{15}N -enriched nitrate. In contrast, deep-water gradients are likely to be subtler, reflecting the basin's relatively stable physical structure at depth and the combined influences of remineralization, mixing, and deep-water formation (Pavlidou et al., 2020; Chaikalis et al., 2021; Tecchio et al., 2013).

Despite these expected gradients, no basin-scale isoscapes have yet been developed for the Mediterranean Sea. This reflects both the lack of spatially representative reference datasets for statistical modeling and the challenges of mechanistically modeling semi-enclosed basins with

Table 1

Summary of key stable isotope applications in the Mediterranean Sea that would benefit from spatialization through isoscapes, encompassing biogeochemical, ecological, and applied management contexts. Isoscapes provide spatially explicit baseline information that supports consistent, basin-scale interpretation of isotopic data and facilitates the integration of local studies. Where available, examples of isoscape-based analyses are provided.

Biogeochemistry	Trophic Ecology	Spatial/ Movement Ecology	Applied Ecosystem Management and Forensics
Constraining the balance of N_2 fixation vs terrestrial/anthropogenic nitrogen inputs in the Mediterranean Sea (Pantoja et al., 2002; Koppelman et al., 2003; Cresson et al., 2016; Liénart et al., 2017; Wald et al., 2025)	Establishing isotopic baselines for spatial inference of trophic geography and structure (Jennings and van der Molen, 2015; Bird et al., 2018; Pethybridge et al., 2018a; Arnoldi et al., 2023) and for the parameterization of isotope mixing models, niche analyses, and assessments of trophic interactions (Koppelman et al., 2009; Chouvelon et al., 2012; Mancinelli et al., 2013; Reddin et al., 2018; MacKenzie et al., 2019; Matich et al., 2021; Merquiol et al., 2023)	Providing natural georeferenced tags to complement artificial tagging (Trueman et al., 2017; Trueman and St John Glew, 2019; McMahon et al., 2013); isoscapes remove the need for directed georeferenced sampling of target species in all possible origin regions (Wunder, 2010; Ma et al., 2020)	Contributing to monitoring frameworks (e.g., MSFD; Berto et al., 2025; Fanelli et al., 2022; Breton et al., 2023; Liénart et al., 2025; Protapapa et al., 2026)
Testing and refining isotope-enabled biogeochemical models (Schmittner et al., 2013; Somes et al., 2010) to improve reconstructions of past (Schmittner and Somes, 2016; Somes et al., 2017; Buchanan et al., 2019b) and projections of future nutrient dynamics (Buchanan et al., 2019a, 2022)	Accounting for isotopic variation in food web models introduced by mobile consumers transporting organic matter across habitats or trophic levels (Trebilco et al., 2016; Travers-Trolet et al., 2019)	Coupling with agent-based models (Trueman et al., 2019; Carpenter-Kling et al., 2019), tagging data (Vander Zanden et al., 2015; Jaeger et al., 2010; St John Glew et al., 2018) or oceanographic models (Arcioni et al., 2025) to improve movement reconstruction accuracy and identify feeding hotspots	Optimizing placement and evaluating effectiveness of MPAs (McCook et al., 2009; Vizzini and Mazzola, 2009; Foley et al., 2010)
Evaluating biogeochemical consequences of deep-water mass formation and its temporal/spatial variability	Informing and testing ecosystem models, facilitating development of isotope-enabled food web models (Jennings et al., 2008; Navarro et al., 2011; Pacella et al., 2013; Pethybridge et al., 2018b;	Quantifying functional connectivity (McCauley et al., 2012; Shipley et al., 2023)	Testing claims of geographic origin of seafood and marine products (e.g., combatting IUU fishing and food fraud, and enhancing regional product protection; Slessor and Trueman, 2021; Cusa et al., 2022)

(continued on next page)

Table 1 (continued)

Biogeochemistry	Trophic Ecology	Spatial/ Movement Ecology	Applied Ecosystem Management and Forensics
Quantifying the proportional contributions of different organic matter sources, including terrestrial and riverine inputs, particularly across coastal-offshore and land-sea gradients (Liénart et al., 2017; Tesi et al., 2007; Sanchez-Vidal et al., 2009; Pedrosa-Pàmies et al., 2015; Rumolo et al., 2015; Tesi et al., 2013; Ferchiche et al., 2025)	McCormack et al., 2019) Tracing nutrient fluxes through food webs across habitats and depth layers, including land-sea interfaces (Darnaude et al., 2004; Darnaude, 2005; Harmelin-Vivien et al., 2008), vertical transfers (Polunin et al., 2001; Fanelli et al., 2013, 2009, 2011; Papiol et al., 2013), and benthic-pelagic coupling (Valls et al., 2014a; Coppari et al., 2020; Masià et al., 2024; Cobain et al., unpubl results)	Identifying routes for pathogen transport via migratory species	Supporting isotope applications in fisheries management (Jennings et al., 2002; Pinnegar et al., 2003; Rumolo et al., 2018; Agnetta et al., 2024)
Tracing the entry and dispersion of anthropogenic nutrient pollutants (Costanzo et al., 2001; Vizzini and Mazzola, 2004; Rumolo et al., 2011; Jona-Lasinio et al., 2015; Rossi et al., 2018; Signa et al., 2020; Thibault et al., 2020)		Enabling retrospective analyses from archived biological materials (Trueman et al., 2019; MacKenzie et al., 2011)	Supporting interpretations of fundamental ecological data for data-deficient and high-conservation-concern species (turtles, seabirds, cetaceans, elasmobranchs; non-lethal sampling; Vander Zanden et al., 2015; St John Glew et al., 2019a; Bird et al., 2018; Arcioni et al., 2025; Di Lorenzo et al., 2020; Smith et al., 2021)
		Investigate spatial aspects and comparisons of biological invasions (Hungate et al., 2016; McCue et al., 2020; Bernardi et al., 2024)	

complex circulation dynamics (Magozzi et al., 2017; Somes et al., 2010). However, the extensive collection of local and regional isotope datasets provides an opportunity to construct a spatially broad, albeit heterogeneous, compiled reference dataset. Here, we leverage this opportunity and use the Mediterranean Sea as a test case to evaluate the feasibility of developing broad-scale geographic baseline isoscapes from opportunistically compiled data. We assembled a comprehensive reference dataset integrating available $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ records from diverse sample types representing low-trophic-level compartments of food webs primarily fueled by epipelagic production. We then applied INLA models to test

whether broad-scale isotopic gradients can be resolved from such data, and to assess the extent to which spatial structure can be distinguished from non-spatial sources of variation. Our aim is therefore not only to develop a Mediterranean isoscape, but also to evaluate how sampling design influences the ability to recover spatial patterns in isotopic data. In particular, we focus on cases where sampling is not structured to capture spatial variability, and where additional heterogeneity in sampling conditions may further obscure underlying spatial signals. Ultimately, our goal is to identify the conditions under which isoscape models can provide robust spatial inference, and to inform future sampling strategies in the Mediterranean Sea and other complex marine systems.

2. Materials and methods

2.1. Data compilation

To develop isoscapes, we compiled all available georeferenced $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ data from published and unpublished studies in the Mediterranean region. During data compilation, it became clear that the resulting dataset would unevenly be distributed in space and heterogeneous with respect to sample types, metadata, and sampling conditions, and therefore not appropriately structured relative to expected underlying sources of isotopic variation. Constructing a usable reference dataset under these conditions requires a tradeoff between restricting inclusion to well-documented samples with standardized metadata, and maximizing spatial coverage while accepting greater uncertainty (St John Glew et al., 2019b). As isoscape models require geographically extensive datasets to resolve spatial gradients, we opted for the latter and prioritized spatial coverage.

The compiled dataset included a broad range of sample types, each functionally representative – to varying degrees – of low-trophic-level compartments of marine food webs primarily sustained by epipelagic or pelagic production (Tables S1 and S2). Given inconsistencies in metadata structure, completeness, and granularity across sources, we classified samples into six broad ecological categories: picoplankton, phytoplankton, particulate organic matter (POM), zooplankton, suprabenthos, and sediment, following study-specific categorizations when available. Our category ‘picoplankton’ corresponded to the $< 20\ \mu\text{m}$ fraction, ‘phytoplankton’ to identified phytoplankton and/or the $20\text{--}200\ \mu\text{m}$ fraction, and ‘POM’ to samples described as bulk particulate organic material or broad size fractions ($< 200\ \mu\text{m}$), with distinction between suspended and sinking material retained where available. The ‘zooplankton’ category included bulk, size-fractionated, or taxonomically resolved samples depending on data availability, with size and taxonomic information retained where provided. ‘Suprabenthos’ included assemblages or taxa consistent with suprabenthic ecology, such as caridean and penaeoid decapods, amphipods, isopods, and mysids. ‘Sediment’ samples corresponded to surface sediments or upper sedimentary layers (typically $\leq 0.5\ \text{cm}$). More detailed category descriptions, size classes, and taxonomic levels are provided in Tables S1 and S2. Sample categories were defined to maximize within-group comparability and expected between-group differences while retaining sufficient sample sizes for modeling. This classification provided a pragmatic framework for partitioning variance, enabling the full dataset to be used to model spatial isotope structure while explicitly accounting for systematic differences among categories. We included sediment samples as a category despite remineralization processes, as surface sediment organic matter retains a largely epipelagic origin (Pedrosa-Pàmies et al., 2015; Rumolo et al., 2015; Struck et al., 2001). Similarly, suprabenthic organisms remain closely linked to pelagic production despite interacting with benthic systems (Polunin et al., 2001; Fanelli et al., 2009; Madurell et al., 2008). Only minimal exclusion criteria were applied *a priori*. We excluded samples that did not represent low-trophic-level compartments fueled by pelagic production, including riverine particulate organic matter, sediment cores (affected

by diagenetic alteration), strictly benthic infaunal organisms, and higher-trophic-level consumers with carnivorous diets (e.g., cephalopods, teleosts, and elasmobranchs). Individual baseline samples with $\delta^{13}\text{C}$ and/or $\delta^{15}\text{N}$ values outside plausible bulk marine environmental ranges (-40 to -10‰ and -5 to 15‰ for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$, respectively) were excluded. Samples from very nearshore locations were excluded during model construction, as environmental covariate data were unavailable at those sites given the spatial extent and resolution of the covariate layers (Table S3).

2.2. Environmental covariates

Environmental variables previously shown to correlate with $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ values (e.g., MacKenzie et al., 2014; St John Glew et al., 2019b; Espinasse et al., 2022; Jennings and Warr, 2003; Barnes et al., 2009) and varying across broad spatial gradients in the Mediterranean Sea were selected for the models. Environmental data encompassing both physical and biogeochemical variables were obtained from the Copernicus Monitoring Environment Marine Service (Copernicus Marine Service (CMEMS), 2025), with monthly averages available at $0.042^\circ \times 0.042^\circ$ spatial resolution for the period 2009–2019. Physical variables – sea surface temperature (SST), bottom temperature (BT), sea surface salinity (Sal), mixed layer depth (MLD), and bathymetry (Bathy) – were extracted from the Mediterranean Sea Physics Reanalysis product (MEDITERRANEAN_SEA_MULTITYEAR_PHY_006_004; Escudier et al., 2021). Biogeochemical variables – chlorophyll *a* concentration (Chl), phytoplankton biomass (phytoC), primary production (PP), and concentrations of nitrate (NO_3^-), phosphate (PO_4^{3-}), ammonium (NH_4^+), and dissolved inorganic carbon (DIC) – were obtained from the coupled Mediterranean Sea Biogeochemistry Reanalysis model (MEDITERRANEAN_SEA_MULTITYEAR_BGC_006_008; Cossarini et al., 2021).

We used model-derived climatologies rather than observational (e.g., satellite) datasets for three main reasons. First, satellite products are unavailable for several variables relevant to isotope dynamics (e.g., DIC, nutrients, bottom temperature). Second, model outputs provide internally consistent fields for both physical and biogeochemical variables at matching spatial and temporal resolutions. Third, the CMEMS Reanalysis products used here have been validated against extensive *in situ*, satellite, and climatological observations and show good performance at basin scales (Escudier et al., 2021; Cossarini et al., 2021). For example, physical variables show strong agreement with observations (e.g., RMSD ca. 0.55°C for SST; ca. 0.17 – 0.18 PSU for Sal), while biogeochemical variables effectively reproduce basin-scale distributions and variability (e.g., RMSD ca. 0.04 mg m^{-3} for Chl in open waters; ca. 0.07 – 0.19 mg m^{-3} for NO_3^- across depth layers). Model-derived climatologies therefore capture large-scale spatial gradients and vertical structure across the Mediterranean Sea, although uncertainties increase at finer spatial and temporal scales. This approach ensures spatial and temporal coherence among covariates while maintaining ecological relevance across a broad suite of variables. For integrated surface values, monthly values for each variable were averaged across depths shallower than 40 m , corresponding to the lower limit of the seasonal thermocline in the Mediterranean Sea (Hecht et al., 1988; Benincasa et al., 2024). Monthly climatologies were then computed across years (2009–2019), from which annual averages were derived.

In addition to environmental variables, we included the spatially derived least-cost distance to coast (dCoast) as an additional covariate. This was computed from each grid cell to the nearest coastline using a binary cost surface, where movement through sea cells had low and equal cost (conductance = 1), while land was treated as an impassible barrier (conductance = 0). The cost surface was generated from a raster at the same resolution as the modeled climatologies, and distance calculations were performed using the *gdistance* R package (van Etten, 2017). This metric captures the complex coastal and island geography of the Mediterranean Sea and provides a more realistic representation of marine connectivity than straight-line (Euclidean) distance.

For spatial prediction, we created a regular 50 km resolution grid based on a custom Lambert Azimuthal Equal-Area (LAEA) projection centered on the Mediterranean Sea, covering the domain from -4° to 39.5° longitude and from 27.5° to 44.5° latitude. Covariate values were extracted at each grid point. To avoid extrapolation beyond the range of observed environmental conditions, we excluded prediction points with missing covariate values or with values falling outside 5% of the observed range at sample locations. All covariates were mean-centered prior to modeling (see Fig. S5 for maps of individual environmental variables).

2.3. Sample variograms

To independently assess spatial autocorrelation in the reference samples, we computed empirical semivariograms for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ values (Zuur et al., 2017). Instead of Euclidean distances, we used least-cost distances between sample locations, calculated using the same approach as for dCoast described above. Semivariograms were generated using 100 evenly spaced distance bins from 0 to 4000 km , retaining only bins containing ≥ 500 sample pairs to ensure robust estimates.

To compare basin-scale spatial autocorrelation patterns with those in areas characterized by denser and more spatially uniform sampling, we also computed semivariograms for three subregions: the Adriatic Sea, Gulf of Lion, and Balearic Sea. Bin widths were adjusted to 22.5 , 6.25 , and 8.75 km , respectively, to reflect the smaller spatial extent of these subregions while maintaining sufficient sample support (≥ 100 pairs per bin).

While INLA-based models explicitly account for non-spatial, category-level differences via random effects, semivariograms were computed using all reference samples pooled across categories. As a result, estimates of spatial autocorrelation may be confounded by between-category variation. Category-specific semivariograms were not feasible due to limited spatial coverage within individual categories. Consequently, semivariograms reflect both spatial structure and the sampling structure of categories. Because categories differ in spatial clustering and coverage, short-distance bins may be dominated by within-category comparisons, whereas larger distance bins may increasingly include between-category comparisons.

2.4. INLA models

To construct isoscape models, we used a Bayesian hierarchical spatial modeling framework implemented through the INLA method via the *R-INLA* package (R-INLA, 2025; Rue et al., 2009). This approach allows environmental covariates to be included as linear fixed effects, and both spatial (via a latent Gaussian Markov Random Field; hereafter spatial field) and nominally non-spatial (e.g., categorical groupings) factors are modeled as random effects. The application of INLA to isoscape modeling is described in detail by St John Glew et al. (St John Glew et al., 2019b); here we summarize only the elements specific to this study.

INLA enables flexible modeling of spatial fields using a Delaunay triangulated mesh that accounts for boundary effects (i.e., land; Fig. S7). Observation locations served as the initial mesh vertices, with additional vertices added to optimize triangle size while maintaining higher resolution in areas with denser sampling. To capture broad-scale spatial structure but avoid interpolation artefacts, the maximum triangle edge length was set to 100 km .

Each isotope value was modeled as a function of covariates (where included), sample category, and a spatial field (Eq. (1)):

$$Y_i \sim \text{Intercept} + \beta_i X_i + f(U_i) + f(W_i) + \varepsilon_i$$

$$U_i \sim N(0, \sigma_{\text{category}}^2)$$

$$W_i \sim N(0, \Omega)$$

$$\varepsilon_i \sim N(0, \sigma^2) \quad (1)$$

where Y_i is the isotope value ($\delta^{15}\text{N}$ or $\delta^{13}\text{C}$) at location i , X_i is a vector of covariates with associated coefficients β_i , U_i is the category random effect, W_i represents the spatial field (with Matérn covariance Ω), and ε_i is the independent residual.

In addition to linear fixed effects, covariates were also modeled as smooth terms. Smooth terms were constructed as second-order basis splines using a 1D mesh and corresponding projection (A) matrix, specifying nine equally distributed knots to capture potential non-linear patterns while minimizing the risk of overfitting. Smooth terms were preferred over interaction terms due to their flexibility, which improves predictive performance and facilitates the emergence of non-linear relationships. Non-informative default priors were applied throughout.

We tested a suite of INLA models combining linear or smooth covariates, sample category, and spatial structure, along with null models that included only spatial and category random effects. Model selection was based primarily on Watanabe-Akaike Information Criteria (WAIC), complemented by evaluation of predictive performance (Pearson correlation coefficient, ρ , between observed and fitted values), and posterior estimates of the spatial field's range (spatial autocorrelation) and variance. Using the selected models, we predicted isotope values at the mesh vertices. These predictions were interpolated across mesh triangles using a second-order (quadratic) basis function via Bayesian kriging to generate spatially continuous isoscapes. For each grid cell, we extracted mean and variance predictions to map the expected isotopic composition and associated uncertainty of low-trophic-level compartments of marine food webs primarily fueled by epipelagic production across the Mediterranean Sea.

2.4.1. Covariate and random effect structure

Preliminary model testing was conducted to identify the most appropriate covariate set and random-effect structure for subsequent comparisons among null, linear, and smooth models, and for isoscape prediction. For covariate selection, environmental variables we first assessed for collinearity using pairwise correlations and variance inflation factors (VIFs). Collinear variables were then summarized using principal component analysis (PCA) based on values extracted at each sample location, including temperature (SST), productivity (Chl, phytoC, and PP), and nutrients (NO_3^- , PO_4^{3-} , and NH_4^+). The first principal component (PC1) explained > 80% of the variance and represented a broad west-to-east gradient of increasing temperature and decreasing nutrients and productivity. PC2 explained only 7–8% (for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$, respectively) of the variance and primarily captured finer-scale nutrient contrasts (see Fig. S6 for PCA of full set of environmental variables). Despite its limited contribution to total variance, PC2 was retained for testing to assess potential improvement in model performance. We also compared alternative variables among other collinear pairs (DIC vs Sal, Bathy vs dCoast). Non-collinear variables (BT, MLD) and PC1 were retained in all models. The final covariate set used in model comparisons included BT, MLD, PC1, DIC, and Bathy.

To evaluate random-effect structure, we tested alternative groupings of isotopically similar or ecologically related categories (e.g., combining pico-nanoplankton and POM, and phytoplankton and POM based on expected partial overlap between size fractions; and combining pico-nanoplankton and sediment based on observed $\delta^{13}\text{C}$ similarity). We also explored reduced dataset configurations (e.g., excluding suprabenthos or retaining only POM without a category effect). In addition, we tested the inclusion of season and year as additional random effects using both crossed (Category + Season; Category + Year) and nested (Season within Category) structures in restricted datasets with complete temporal metadata. Across all tests, these alternative structures did not improve predictive performance, spatial structure, or interpretability. For completeness, Category + Season models are reported for comparison. Subsequent analyses therefore focus on the most parsimonious and

informative structure, including category as the sole random effect with all six category levels and the full dataset.

3. Results

3.1. The Mediterranean Sea reference dataset

The compiled dataset comprised 5,428 samples from 595 locations for $\delta^{15}\text{N}$ and 6,034 samples from 610 locations for $\delta^{13}\text{C}$ across the Mediterranean Sea (Tables S1 and S2). Despite its size, spatial coverage was highly uneven. Sampling density was highest in the Gulf of Lion and Balearic Sea, moderate in the Adriatic, Ionian, and Aegean-Cretan Seas, and sparse in the southern and southeastern Mediterranean, particularly in the southern Levantine Basin (Fig. 1 and S2A, B).

The distribution of sample categories also varied spatially (Fig. 1 and S2C, D). POM and zooplankton accounted for over 75% of samples in most areas, although their relative contributions differed regionally. Zooplankton dominated datasets from the Adriatic (58–65% for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$), Ligurian (60–63% for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$), and Balearic Seas (52–55% for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$), whereas POM was most prevalent in the Aegean Sea (91–95% for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$) and eastern Mediterranean (0.53–64% for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$). Pico-nanoplankton, phytoplankton, suprabenthos, and sediment were less frequently sampled and more unevenly distributed. Sediment samples were more common in the Tyrrhenian (16–17% for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$), Adriatic (19–27% for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$), and Ionian Seas (3–19% for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$); pico-nanoplankton and phytoplankton were concentrated in the Ligurian Sea (17–18% for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$); and suprabenthos was represented only in the western Mediterranean (2–3% for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$) and Balearic (20–22% for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) and Ionian Seas (22–30% for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$).

Vertical and seasonal sampling structure varied among sea areas and categories. Approximately 61–66% of samples (for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$, respectively) were collected from epipelagic depths (< 200 m), primarily comprising pico-nanoplankton, phytoplankton, POM, and zooplankton (Fig. S3). Deeper samples contributed 21–23% ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) from mesopelagic, 8–10% ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) from bathypelagic, and < 1% (both isotopes) from hadal depths, and were mainly represented by POM, zooplankton, and sediment in the Balearic, Adriatic, Ionian and Aegean Seas, as well as in the eastern Mediterranean. Suprabenthos samples originated primarily from mesopelagic depths.

Total sampling period spanned from October 1993 to March 2022 for $\delta^{15}\text{N}$ and from July 1995 to March 2022 for $\delta^{13}\text{C}$. Seasonal coverage was uneven across sea areas and categories, with many category x season combinations underrepresented or absent (Fig. S4). Metadata on depth, year, and season were frequently incomplete or inconsistently reported. Together, these patterns highlight the uneven and heterogeneous structure of the reference dataset.

3.2. Isotopic variability in the Mediterranean Sea

Values of $\delta^{15}\text{N}$ in the reference dataset had a mean $\pm$ SD of $3.6 \pm 2.3\text{‰}$, with 95% of values ranging from -1.1 to 8.4‰ . Values of $\delta^{13}\text{C}$ had a mean $\pm$ SD of $-22.6 \pm 2.7\text{‰}$, spanning a 95% interval from -27.3 to -15.7‰ . Within-category variability (red diagonal values in Table S4) ranged between 0.4 and 1.9‰ (mean SD) for $\delta^{15}\text{N}$ and between 0.2 and 1.7‰ for $\delta^{13}\text{C}$. Sediment exhibited the lowest variability for both isotopes, whereas pico-nanoplankton and suprabenthos showed the highest variability for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$, respectively. Overall, $\delta^{15}\text{N}$ values showed slightly greater variability than $\delta^{13}\text{C}$ values, except in phytoplankton and suprabenthos.

Between-category mean absolute differences (black off-diagonal values in Table S4) ranged from 0.3 to 3.5‰ for $\delta^{15}\text{N}$ and from 0.1 to 4.1‰ for $\delta^{13}\text{C}$. On average, sediment was enriched in ^{15}N by 0.3‰ relative to POM (the smallest difference), while suprabenthos was enriched by 3.5‰ relative to sediment (the largest). Zooplankton was also ^{15}N -enriched by 3.2‰ compared to pico-nanoplankton but

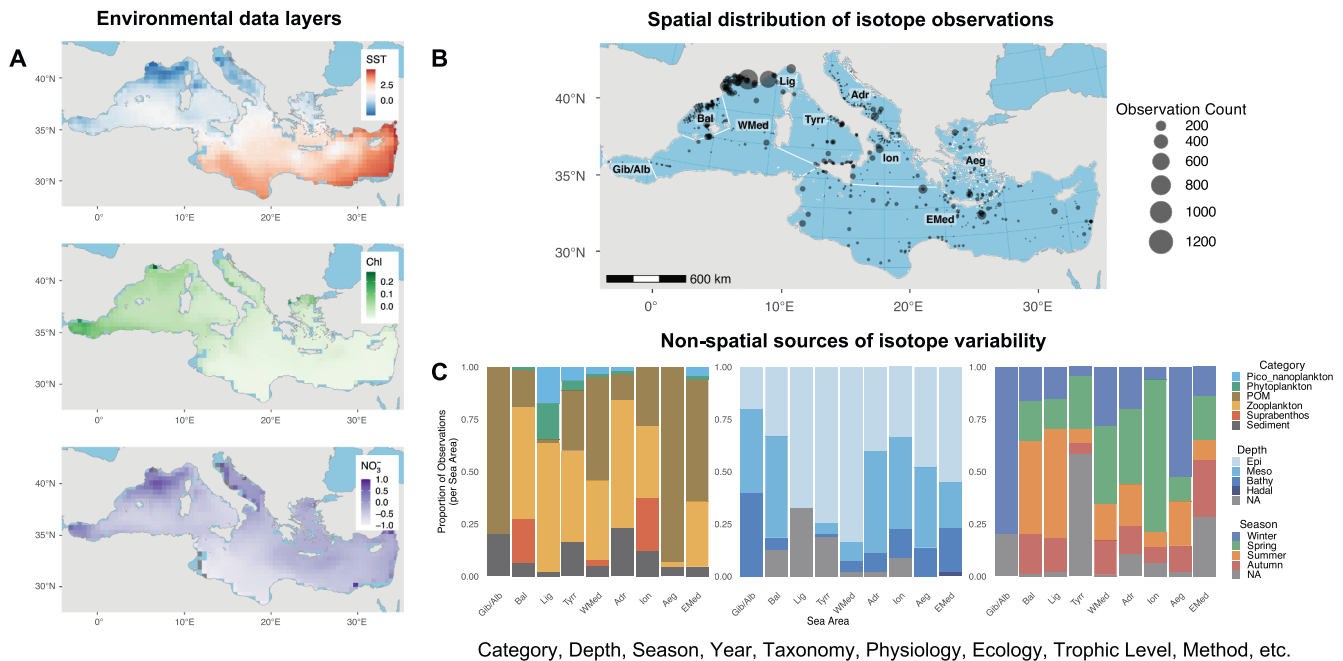


Fig. 1. Conceptual overview of the main data inputs for INLA models: environmental data layers (A), spatial distribution of isotope observations (B), and non-spatial sources of isotope variability (C). Panel A shows three example environmental layers – sea surface temperature (SST [$^{\circ}\text{C}$]), chlorophyll *a* concentration (Chl [mg m^{-3}]), and nitrate concentration (NO_3^- [mmol m^{-3}]) – illustrating basin-scale gradients in temperature, productivity, and nutrient availability across the the Mediterranean Sea. Covariates are shown as mean-centered annual averages of multi-year climatologies. Additional environmental variables and principal component analyses are presented in Fig. S5 and S6. Panel B shows sampling locations of isotope observations, with circle size proportional to the number of observations per location. International Hydrographic Organization (IHO) sea area boundaries are indicated (Gib/Alb = Strait of Gibraltar/Alboran Sea, Bal = Balearic Sea, Lig = Ligurian Sea, Tyrr = Tyrrhenian Sea, NWMed = North-Western Mediterranean Sea, Adr = Adriatic Sea, Ion = Ionian Sea, Aeg = Aegean Sea, EMed = Eastern Mediterranean Sea; see Fig. S1). Observations are pooled across $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$; isotope-specific distributions are shown in Fig. S2A, B. Panel C shows the proportional representation of category, depth zone, and season across IHO sea areas, based on pooled observations. Isotope-specific proportions are provided in Fig. S2C, D (category), S3 (depth), and S4 (season). Additional sources of isotopic variability include year, taxonomy, biological and ecological traits, trophic structure, and sampling and analytical methods.

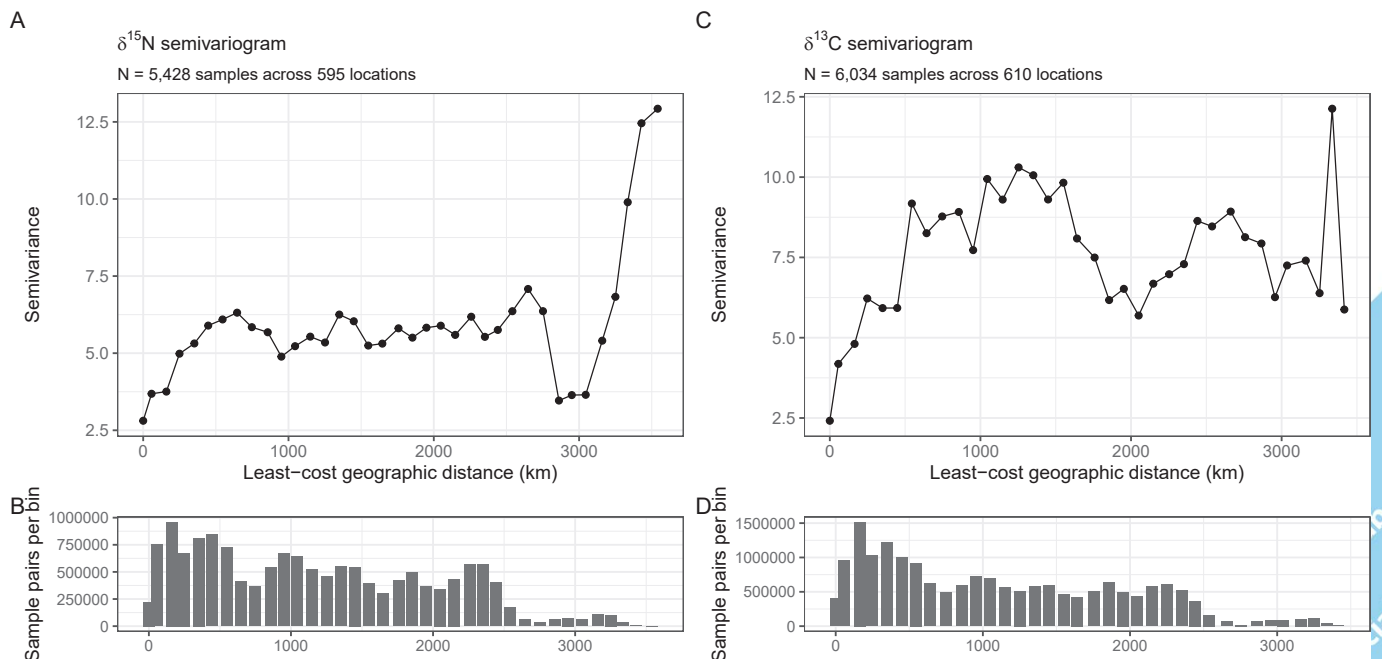


Fig. 2. Sample semivariograms of $\delta^{15}\text{N}$ (A) and $\delta^{13}\text{C}$ (C) reference data computed using least-cost geographic distance across the Mediterranean Sea. Panels B and D show the number of sample pairs per distance bin; support declines substantially beyond ca. 2500 km.

unexpectedly depleted by 2.7‰ relative to phytoplankton. For $\delta^{13}\text{C}$, the smallest difference (0.1‰) was between zooplankton and sediment, with zooplankton slightly enriched in ^{13}C relative to sediment. The largest difference (4.1‰) occurred between sediment and phytoplankton, with phytoplankton enriched relative to sediment. Zooplankton was ^{13}C -enriched relative to both POM (2.2‰) and picoplankton (2.1‰). Overall, between-category differences were often comparable in magnitude to within-category variability.

3.3. Sample variograms

Least-cost distance semivariograms showed that category-pooled variance in both $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ values increased sharply at small spatial scales (within a few hundred kilometers; Fig. 2). For $\delta^{15}\text{N}$, semivariance increased steeply within the first 250 km, continued to increase until ca. 600 km, and then fluctuated inconsistently thereafter (Fig. 2A). The $\delta^{13}\text{C}$ semivariogram showed a similarly steep increase within the first 250 km until ca. 750 km, followed by a noisy plateau and a marked decrease between ca. 1,600 and 2,000 km (Fig. 2B). Spatial autocorrelation in both isotopes was therefore limited to short distances – on the order of a few hundred kilometers or less – beyond which isotope values became statistically independent. Subregional semivariograms exhibited similarly weak or shorter-range spatial autocorrelation than at the basin scale (Fig. S8), with some regions showing no detectable autocorrelation. This indicates that small-scale variability persisted even in regions with denser and more spatially uniform sampling.

Because semivariograms were computed across pooled categories, these patterns reflect both spatial structure and differences in sampling among categories. Categories differed markedly in spatial support, with mean nearest-neighbor least-cost distance ranging from 12.5–12.7 km (for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$, respectively) for suprabenthos to 138.5–144.5 km (for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) for pico-nanoplankton. As a result, short-distance bins are dominated by comparisons among spatially clustered samples within categories, whereas comparisons at larger distances increasingly include samples from more spatially dispersed categories and between-category contrasts. Consequently, semivariance reflects a mixture of spatial and category-related variability that varies with distance.

3.4. INLA models

3.4.1. Model comparison

Comparisons of spatial vs non-spatial models showed that including a spatial field substantially improved both model fit (WAIC) and predictive accuracy (ρ) for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$. For $\delta^{15}\text{N}$ null models, WAIC decreased from 23,453.29 in the non-spatial model (M0) to 21,945.75 in its spatial counterpart (M0.SF), while ρ increased from 0.43 to 0.66. For $\delta^{13}\text{C}$, WAIC decreased from 26,490.21 (M0) to 24,215.33 (M0.SF), with ρ increasing from 0.59 to 0.79 (Table 2). Similar improvements were observed when comparing non-spatial covariate-based models (M1lin, M1sm) with their spatial counterparts (M1lin.SF, M1sm.SF; Table 2).

Within the spatial model set, the inclusion of environmental covariates improved model fit for $\delta^{15}\text{N}$ relative to the null spatial model when included as smooth terms, but not as linear fixed effects. For $\delta^{13}\text{C}$, the inclusion of covariates – either as linear or smooth terms – did not improve model fit relative to the null spatial model. For $\delta^{15}\text{N}$, the lowest WAIC values were obtained for the smooth covariate model (M1sm.SF: 21,873.78), compared to 21,945.75 for the null spatial model (M0.SF) and 21,948.82 for the linear covariate model (M1lin.SF). In contrast, for $\delta^{13}\text{C}$ the lowest WAIC values were obtained for the null spatial model (M0.SF: 24,215.33), with higher values for models including covariates (M1lin.SF: 24,257.29; M1sm.SF: 24,264.02). Based on WAIC, the best-performing models were therefore M1sm.SF for $\delta^{15}\text{N}$ and M0.SF for $\delta^{13}\text{C}$. Predictive accuracy, however, remained nearly unchanged across models ($\delta^{15}\text{N}$: $\rho = 0.66$ – 0.67 ; $\delta^{13}\text{C}$: $\rho = 0.79$; Table 2), indicating that improvements in WAIC were not accompanied by gains in predictive performance. Thus, while smooth covariate terms improved model fit for $\delta^{15}\text{N}$, they did not enhance predictive accuracy despite increased model complexity. Consistent with this, smooth covariate effects were weak and poorly constrained, with no clear or consistent trends (Fig. S9).

3.4.2. Spatial field random effect

Estimated spatial fields were characterized by short ranges and high variances across models (Fig. 3 and S10), indicating that spatial isotope variability in the dataset non-attributable to environmental covariates (when present in the models) was dominated by strong small-scale structure. For $\delta^{15}\text{N}$, the smooth covariate model (M1sm.SF) yielded a mean spatial field range of 65.4 km (95% credible interval: 48.4–85.8 km). Similar ranges were obtained for the null spatial model (M0.SF: 75.5 km; 57.8–99.1 km) and the linear covariate model (M1lin.SF: 72.7

Table 2

Model comparison results for the $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ models. Models tested the inclusion of a spatial field as a random effect and environmental covariates as either linear fixed effects or smooth terms. All models included category as a non-spatial random effect. Model fit was evaluated using the Watanabe-Akaike Information Criteria (WAIC), and predictive accuracy was assessed from the Pearson correlation coefficient (ρ) between observed and fitted values. The effective number of parameters (Eff. n parms) and most influential covariates (i.e., significant linear fixed effects or smooth terms with the largest attributed variance) are also reported. For correlations, the t , ρ , 95% confidence interval, and degrees of freedom (df) are shown. All correlations were statistically significant (p -value < 0.001). Models are listed in order of increasing WAIC for each isotope, with ΔWAIC reported relative to the best model.

Model	Model formula	WAIC	ΔWAIC	Eff. n parms	t	ρ	95% confidence interval	df	Significant covariates
$\delta^{15}\text{N}$									
M1sm.SF	Smooth + f(Category) + SF	21,873.78	na	251.73	66.04	0.67	0.65–0.68	5,426	Bathy, DIC, MLD, PC1
M0.SF	f(Category) + SF	21,945.75	71.97	245.63	64.90	0.66	0.65–0.68	5,426	na
M1lin.SF	Linear + f(Category) + SF	21,948.82	75.04	247.40	64.90	0.66	0.65–0.68	5,426	MLD
M1sm	Smooth + f(Category)	22,895.48	1,021.70	42.25	45.08	0.52	0.50–0.54	5,426	All
M1lin	Linear + f(Category)	23,346.84	1,473.06	14.19	36.93	0.45	0.43–0.47	5,426	Bathy, BT, MLD, PC1
M0	f(Category)	23,453.29	1,579.51	9.63	34.93	0.43	0.41–0.45	5,426	na
$\delta^{13}\text{C}$									
M0.SF	f(Category) + SF	24,215.33	na	464.18	99.31	0.79	0.78–0.80	6,032	na
M1lin.SF	Linear + f(Category) + SF	24,257.29	41.96	507.26	99.83	0.79	0.78–0.80	6,032	Bathy, MLD
M1sm.SF	Smooth + f(Category) + SF	24,264.02	48.69	575.72	101.56	0.79	0.78–0.80	6,032	Bathy, DIC, MLD, PC1
M1sm	Smooth + f(Category)	25,717.11	1,501.78	45.76	67.88	0.66	0.64–0.67	6,032	All
M1lin	Linear + f(Category)	26,220.12	2,004.79	17.53	60.57	0.61	0.60–0.63	6,032	Bathy, BT, DIC, PC1
M0	f(Category)	26,490.21	2,274.88	12.27	56.87	0.59	0.57–0.61	6,032	na

Model notation: M0 = null model (no environmental covariates); M1lin = model with linear covariates; M1sm = model with smooth covariates; “.SF” indicates inclusion of a spatial field.

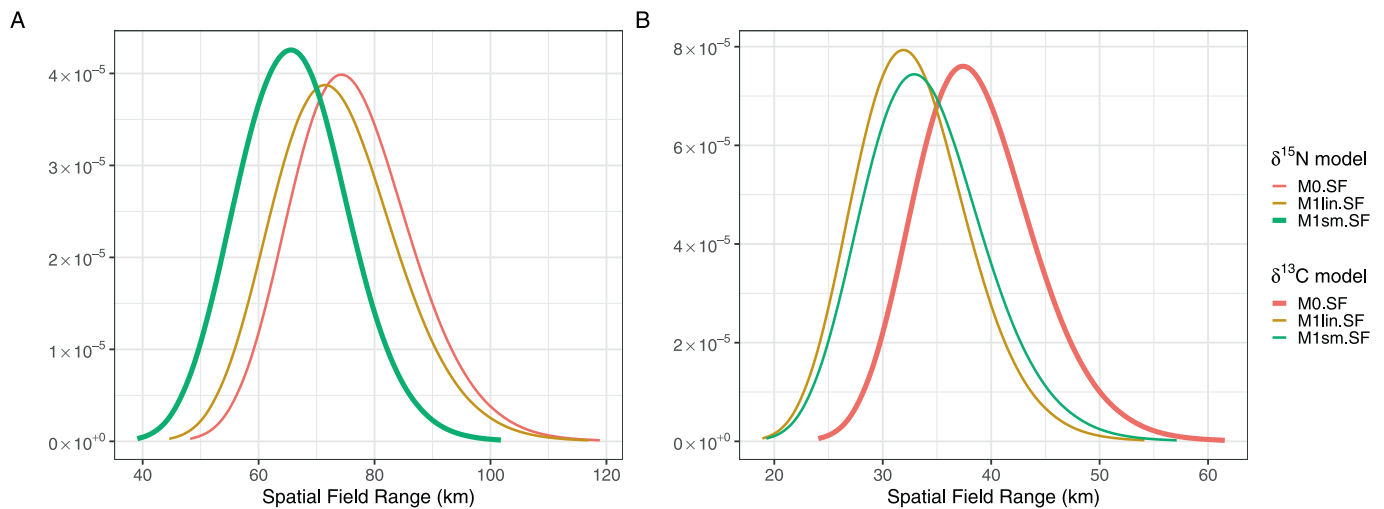


Fig. 3. Posterior probability distributions of the spatial field range parameter for $\delta^{15}\text{N}$ (A) and $\delta^{13}\text{C}$ (B) models. All models include the spatial field and category as random effects. Models shown are: (i) null spatial models without covariates (M0.SF), (ii) models with covariates as linear fixed effects (M1lin.SF), and (iii) models with covariates as smooth terms (M1sm.SF). Thicker lines indicate the best-performing models based on WAIC (M1sm.SF for $\delta^{15}\text{N}$ and M0.SF for $\delta^{13}\text{C}$).

km; 54.4–96.7 km; Fig. 3A). Thus, including covariates resulted in only a marginal reduction in spatial range relative to the null model when modeled as smooth terms, and no reduction when modeled as linear effects. For $\delta^{13}\text{C}$, spatial field ranges were even shorter and similarly unaffected by the inclusion of covariates: M0.SF: 38.3 km (29.0–50.8 km), M1lin.SF: 32.4 km (23.6–44.2 km), and M1sm.SF: 33.6 km (24.2–46.2 km; Fig. 3B). Mean spatial ranges were only slightly greater than half of the maximum mesh edge length (100 km) for $\delta^{15}\text{N}$, and less than half for $\delta^{13}\text{C}$. For reference, the mean nearest-neighbor distance among unique sample locations was 22.2 ± 28.1 km (median 13.2 km) for $\delta^{15}\text{N}$ and 19.0 ± 38.7 km (median 7.0 km) for $\delta^{13}\text{C}$.

3.4.3. Non-spatial random effects

Marginal posterior distributions of the category random effect are shown in Fig. 4. Substantial isotopic variability among categories (ca. 4‰ for both $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$) persisted across models, indicating this variation was largely independent of environmental covariates and continuous spatial structure. For $\delta^{15}\text{N}$, pico-nanoplankton had the lowest values, phytoplankton and suprabenthos the highest, and POM, sediment, and zooplankton were intermediate (Fig. 4A). For $\delta^{13}\text{C}$, POM,

pico-nanoplankton and sediment showed the lowest values, followed by phytoplankton and zooplankton, with suprabenthos highest (Fig. 4B). Posterior distributions overlapped more for $\delta^{15}\text{N}$ than for $\delta^{13}\text{C}$. For $\delta^{15}\text{N}$, strong overlap occurred among POM, sediment, and zooplankton, and between phytoplankton and suprabenthos. In contrast, $\delta^{13}\text{C}$ posteriors showed clearer separation, with overlap mainly among POM, sediment, and pico-nanoplankton, and between phytoplankton and zooplankton, while suprabenthos remained distinct (Fig. 4).

The inclusion of season as an additional random effect, using a reduced dataset (88–89% of samples for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$) with complete seasonal metadata, marginally improved WAIC values for both isotopes ($\delta^{15}\text{N}$: WAIC: Category + Season = 19,092.34 vs Category = 19,189.84; $\delta^{13}\text{C}$: WAIC: Category + Season = 21,129.12 vs Category = 21,167.19). However, predictive accuracy and spatial range remained unchanged (Table S5). Posterior variability among seasons was substantially smaller (ca. 1‰ for $\delta^{15}\text{N}$, 0.5‰ for $\delta^{13}\text{C}$) than among categories, and seasonal posteriors overlapped extensively (Fig. S12), with limited influence on category estimates.

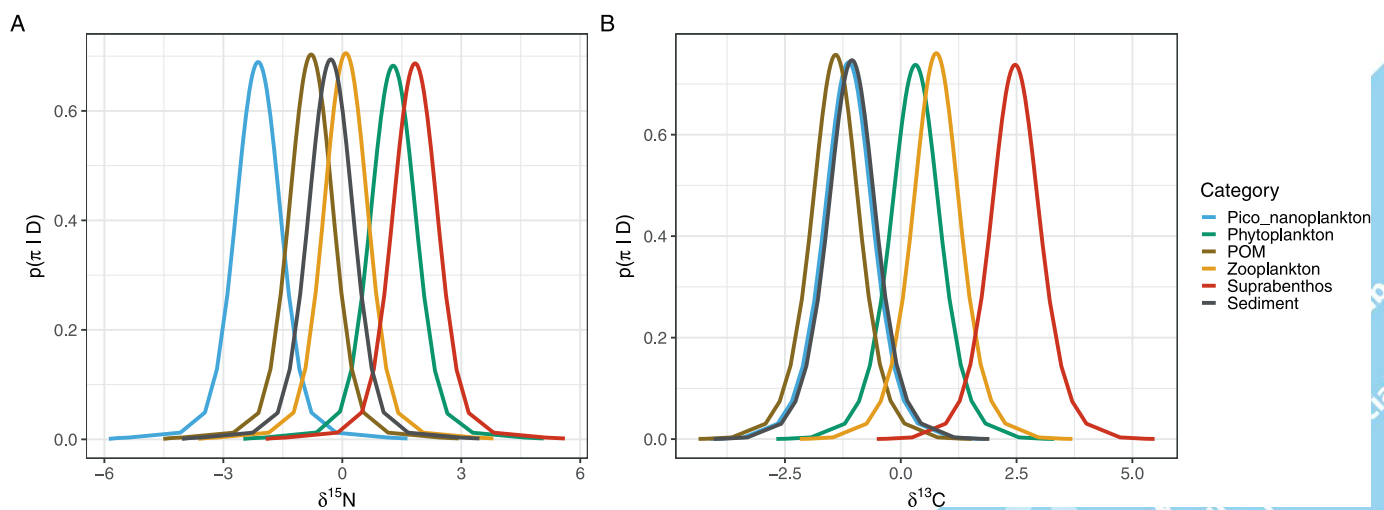


Fig. 4. Marginal posterior distributions of the category random effect for $\delta^{15}\text{N}$ (A) and $\delta^{13}\text{C}$ (B) models. Each distribution ($\pi | D$) represents the probability density of the deviation of a given category from the overall mean isotope value, conditional on the data. Distributions are derived from the best-performing models (M1sm.SF for $\delta^{15}\text{N}$ and M0.SF for $\delta^{13}\text{C}$).

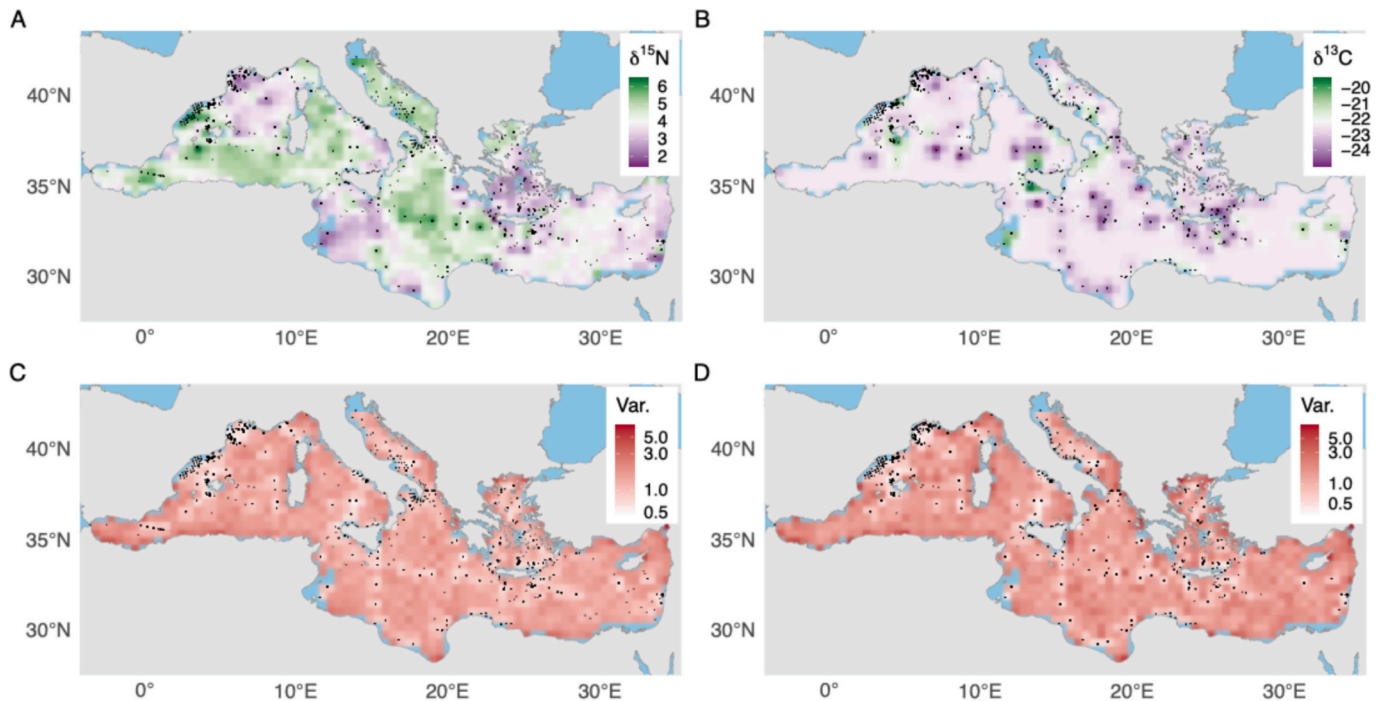


Fig. 5. Predicted isoscape surfaces for $\delta^{15}\text{N}$ (A, C) and $\delta^{13}\text{C}$ (B, D) across the Mediterranean Sea. Panels show posterior means (A, B) and posterior variances (C, D) predicted by the best-performing models (M1sm.SF for $\delta^{15}\text{N}$ and M0.SF for $\delta^{13}\text{C}$). Variances are displayed on a $\log_{10}$ color scale for visualization. Points indicate sampling locations. Predictions represent geographic baseline isoscapes derived from six sample categories representing low-trophic-level food web compartments primarily supported by epipelagic production. Category was included as a non-spatial random effect to account for systematic differences among groups.

3.4.4. Predicted isoscapes

Isoscapes predicted using the best-performing models (M1sm.SF for $\delta^{15}\text{N}$ and M0.SF for $\delta^{13}\text{C}$) showed that predictions were spatially constrained around areas with high sampling density (Fig. 5). In regions farther from sampled locations, predicted values rapidly reverted to the model's global mean: $4.1 \pm 0.8\text{‰}$ (mean $\pm$ SD) for $\delta^{15}\text{N}$ (Fig. 5A) and $-22.5 \pm 0.6\text{‰}$ for $\delta^{13}\text{C}$ (Fig. 5B). Prediction uncertainty was high across the domain, with mean $\pm$ SD predicted variances of $1.5 \pm 0.6\text{‰}$ for $\delta^{15}\text{N}$ (range: 0.5–7.1‰) and $1.6 \pm 0.9\text{‰}$ for $\delta^{13}\text{C}$ (range: 0.4–7.8‰). The spatial distribution of uncertainty closely mirrored that of the predicted means, with lowest uncertainty at sampled locations and consistently high uncertainty across unsampled areas (Fig. 5C, D).

In both isoscapes, mean spatial variation in predicted values (0.8‰ for $\delta^{15}\text{N}$ and 0.6‰ for $\delta^{13}\text{C}$) was smaller than mean posterior uncertainty (1.5‰ and 1.6‰, respectively), indicating that spatial differences were weak relative to model uncertainty. The $\delta^{15}\text{N}$ isoscape showed apparent differences among discrete, well-sampled regions – for example, lower predicted values in the Gulf of Lion and Aegean Sea compared to higher values in the Balearic and Adriatic regions (Fig. 5A). These differences exceeded local prediction uncertainty within well-sampled regions, but did not persist across the distances separating those regions, such that spatial differences were not maintained across poorly sampled areas.

Predictions from models including covariates showed greater apparent structure than those from null spatial models (Fig. S13–S15). However, this structure primarily reflected the spatial distribution of the covariates used for prediction rather than improved recovery of isotope gradients. Relative to null models, predictive accuracy remained unchanged, spatial ranges did not decrease, and uncertainty remained high across the domain. Overall, covariate-based models did not produce more accurate or more spatially resolved predictions than the null spatial models.

4. Discussion

This study represents a first attempt to construct basin-scale isoscapes for the Mediterranean Sea using a large, opportunistically compiled dataset of $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ values from low-trophic-level compartments. Using a Bayesian INLA framework (St John Glew et al., 2019b), we evaluated whether environmental covariates and spatial autocorrelation could resolve broad-scale (hundreds of kilometers) geographic structure in baseline isotope values while accounting for sample heterogeneity through a category random effect.

Our main finding is that spatial variance in the dataset was dominated by small-scale structure for both isotopes (ca. 40 km for $\delta^{13}\text{C}$ and 65 km for $\delta^{15}\text{N}$; Fig. 3), while the inclusion of environmental covariates did not improve predictive performance (Table 2). This contrasts with expectations for systems structured by broad-scale, environmental gradients, where covariates explain large-scale variation and the spatial field captures only weak, localized variance that cannot be attributed to the covariates or other nominally non-spatial random effects. Instead, both null and covariate-based models consistently produced short-range, high-variance spatial fields, indicating that the spatial component primarily fitted strong, localized non-attributable variance, rather than coherent, basin-scale structure.

This pattern reflects a mismatch between the spatial structure of isotopic variation and the sampling design. While isotopic variation is expected to follow environmental gradients, the compiled dataset is not structured to resolve these gradients or to separate spatial from non-spatial sources of variation. Sampling is uneven relative to expected spatial variation, and non-spatial sources of variability (e.g., category, season, year, depth, taxonomy, method) are not only heterogeneous but also spatially clustered. As a result, non-spatial effects become confounded with space, and the spatial field absorbs variability arising from true fine-scale heterogeneity and unresolved, spatially structured non-spatial variation. This explains why spatial models outperformed non-spatial ones (Table 2) despite failing to resolve broad-scale gradients.

Consequently, predicted isoscapes are constrained to short distances and revert to near-uniform mean values outside well-sampled areas, limiting their ability to represent continuous basin-scale patterns. That is, predicted isoscapes primarily reflect the structure of the underlying dataset rather than spatially resolved isotopic baselines. Apparent regional differences – such as lower $\delta^{15}\text{N}$ values in the Gulf of Lion, Liguro-Provençal basin, and southeastern Mediterranean, and higher values in the Catalan-Balearic, Adriatic, and northern Aegean regions – likely reflect real contrasts among discrete, well-sampled regions. However, these differences cannot be reliably extrapolated beyond sampled areas. They are captured as localized features rather than components of a continuous spatial gradient, and their broader spatial extent and connectivity across poorly sampled regions remain unresolved.

4.1. Are there continuous isotopic gradients in the Mediterranean Sea?

Our results indicate that isotopic variation in the compiled dataset in both $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ is dominated by localized structure rather than clear basin-scale gradients. This pattern could arise from two non-mutually exclusive scenarios: broad-scale isotope gradients may be genuinely weak or absent in the Mediterranean Sea, or they exist but are not resolved given the characteristics of the compiled dataset.

Regarding the first scenario, several independent lines of evidence suggest that broad-scale isotope gradients *do* exist in the Mediterranean Sea. Key environmental drivers of $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ – temperature, nutrient availability, productivity, and nitrogen sources – differ markedly between the cooler, more eutrophic western basin and the warmer, oligotrophic eastern basin (Fig. 1 and S5; Coll et al., 2010; Salgado-Hernanz et al., 2019; Reale et al., 2022; Salgado-Hernanz et al., 2022; Álvarez et al., 2023). These variables underpin mechanistic isoscape models at global scales (Schmittner et al., 2013; Magozzi et al., 2017; Somes et al., 2010) and are established predictors in sample-based isoscapes (e.g., St John Glew et al., 2019b; Espinasse et al., 2022). The pronounced longitudinal gradients in these variables provide a strong theoretical basis for expecting corresponding isotope gradients in organic production. Empirical observations further support this expectation. Basin-scale variation in $\delta^{13}\text{C}$ DIC has been documented ($< 0.5\%$; Pierre, 1999; Sisma-Ventura et al., 2016), consistent with productivity and temperature gradients. Regional differences in $\delta^{13}\text{C}$ and/or $\delta^{15}\text{N}$ values have also been reported across a range of taxa and ecosystem compartments, including POM (Chaikalis et al., 2021), crustaceans (Cartes et al., 2014), cephalopods (David Wells et al., 2021), fishes (Andrews et al., 2023), and sea turtle hatchlings (Cardona et al., 2014), as well as at finer spatial scales in coastal fishes (Jennings et al., 1997; Rumolo et al., 2016). Even within our dataset, winter POM samples show consistent, albeit subtle, differences between the western and eastern basins ($\delta^{15}\text{N}$: 0.4‰, $W = 46206$, $p\text{-value} < 0.05$; $\delta^{13}\text{C}$: 1.2‰, $W = 41350$, $p\text{-value} < 0.05$). Similarly, the full $\delta^{15}\text{N}$ dataset (and corresponding isoscape) supports differences among discrete, well-sampled regions, although spatial modeling cannot determine whether these represent continuous gradients or isolated regional signals (Fig. 5). Taken together, these observations indicate that isotopic gradients likely exist at both basin and regional scales in the Mediterranean Sea.

Nonetheless, basin-scale gradients may be relatively weak. For carbon, the longitudinal temperature gradient is modest (ca. 5 °C across the basin; Fig. 1 and S5) compared to the stronger latitudinal contrasts (2–3-fold larger) that drive pronounced $\delta^{13}\text{C}$ isoscapes in larger oceanic systems (Espinasse et al., 2020; St John Glew et al., 2021), though comparable to variation in UK shelf seas (St John Glew et al., 2019b). Similarly, chlorophyll *a* concentrations in the Mediterranean (ca. 0.2 mg m^{-3} ; Fig. 1 and S5) are substantially lower (ca. 1–2 orders of magnitude) than in more productive systems (St John Glew et al., 2019b; Espinasse et al., 2020, 2022; Cobain et al., unpubl results), potentially resulting in weaker isotope gradients. Moreover, environmental controls can act in opposing directions – for example, cooler temperatures in the west tend

to lower phytoplankton $\delta^{13}\text{C}$ values (Rau et al., 1989), whereas higher nutrient availability can elevate it through increased growth rates (Rau et al., 1996, 1997; Laws et al., 1995, 1997). For nitrogen, contrasts in source contributions between the two basins (Pantoja et al., 2002; Koppelman et al., 2003; Struck et al., 2001) may be dampened by overlapping processes (Wald et al., 2025; Mara et al., 2009; Emeis et al., 2010; Krom et al., 2010; Ibellio et al., 2010; Benavides et al., 2016), including N_2 fixation in both basins (Bonnet et al., 2011; Rahav et al., 2013; Ridame et al., 2022) and variable anthropogenic inputs. Basin-scale circulation and mixing further homogenize surface baselines, reducing the isotopic expression of underlying biogeochemical gradients (e.g., Chaikalis et al., 2021; Merquiol et al., 2023). Interannual variability in circulation and Atlantic inflow may additionally disrupt or weaken persistent large-scale patterns (Millot and Taupier-Letage, 2005).

4.2. Dataset limitations and implications for modeling continuous isotopic gradients

While environmental gradients and biogeochemical processes may constrain the strength of isotope patterns, the ability to detect such gradients ultimately depends on how the data are structured relative to those processes. In this study, the compiled dataset was not suitably structured to resolve spatial isotopic variation expected along environmental gradients. Sampling was uneven in space and confounded multiple sources of variability (e.g., category, season, year, depth, taxonomy, method), which were not independently represented across the basin and were, to varying extents, themselves spatially structured. Although hierarchical models are designed to accommodate unbalanced sampling, the co-occurrence of these sources of variability in space limits the ability to separate spatial from non-spatial effects. As a result, spatial and non-spatial sources of variation could not be effectively disentangled.

The key limitation is not simply uneven sampling, but that non-spatial sources of variation are themselves spatially structured. When non-spatial factors are spatially clustered or covary with environmental gradients, they become confounded with space. Under these conditions, non-spatial variability can be expressed as spatial signal, and the spatial field absorbs variance arising from both true geographic structure and unresolved, spatially structured non-spatial effects. Consequently, the spatial component does not represent purely geographic variation, but reflects the combined structure of the data.

The issue is further amplified by the spatial distribution of sampling. Observations are strongly concentrated in coastal environments, while offshore regions – particularly in the southern and eastern Mediterranean – are comparatively underrepresented. Coastal environments are inherently more isotopically variable due to localized biogeochemical inputs (e.g., river plumes, lagoons), short-term temporal fluctuations, complex trophic interactions, and anthropogenic influences (e.g., Bănară et al., 2014; Liénart et al., 2017; Signa et al., 2020; Darnaud et al., 2004; Cresson et al., 2012; Miller et al., 2013; Chen et al., 2023), reducing signal-to-noise ratios in the data. In contrast, offshore regions typically exhibit lower variability and provide the spatial context necessary to resolve basin-scale gradients. This imbalance does not simply reduce coverage, but disproportionately weights model fitting toward highly variable, spatially clustered observations, further increasing the likelihood that non-spatial variability is expressed as spatial structure.

Among non-spatial effects, sample category was the only factor that could be consistently modeled across the full dataset, yet it captured only part of the variability, accounting for ca. 32.8–38.5% (for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$, respectively) of the combined category and residual variance. Substantial differences among categories (ca. 4‰ for both $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$) persisted across models, confirming that category represents an important source of isotopic variation independent of environmental covariates and spatial structure. If unaccounted for, such variability

could bias inference, as differences of this magnitude represent a major portion of the spatial variance typically observed in marine isoscapes. While the category random effect captured some structured differences among groups, posterior distributions often overlapped between ecologically distinct categories or deviated from expected trophic patterns (Fig. 4), particularly for $\delta^{15}\text{N}$, which is typically trophically structured due to substantial trophic discrimination (DeNiro and Epstein, 1978, 1981; Post, 2002). For instance, zooplankton, as consumers, would be expected to exhibit higher $\delta^{15}\text{N}$ than phytoplankton, yet the reverse was observed (Fig. 4A). These inconsistencies likely arise because category distributions covary with location and other unmodeled factors, rather than reflecting purely ecological differences. Consequently, the category effect absorbed both true group-level differences and spatially confounded or non-categorical variability, limiting its interpretability.

Other key sources of variation – particularly season and depth – were likely among the strongest drivers of isotopic variability (e.g., Polunin et al., 2001; Fanelli et al., 2013, 2011; Papiol et al., 2013; Cresson et al., 2012; Savoye et al., 2003; Valls et al., 2014b) but could not be reliably incorporated due to incomplete or inconsistent metadata and their uneven, spatially structured representation. For instance, sampling season was often missing or not comparable across regions with differing climatic and oceanographic regimes (e.g., eastern and western basins) and among sample types with different isotopic incorporation rates (e.g., phytoplankton vs suprabenthic invertebrates). Critically, season and depth were not sufficiently represented across space and categories to support their inclusion as random effects in the full dataset. Including season as an additional crossed random effect in a reduced dataset with complete metadata marginally improved model fit but did not enhance predictive performance or alter spatial predictions (Table S5), indicating that additional model complexity described internal dataset structure rather than resolving underlying processes. Under these conditions, season likely functions in part as a sampling block (e.g., reflecting individual surveys or cruises) rather than representing a coherent temporal signal, as also suggested by the lack of consistent patterns in posterior distributions (Fig. S12). Similarly, depth-related processes – including subsurface production, vertical transport, and remineralization – introduce additional variability, as reworking extends trophic pathways before incorporation into sampled pools (e.g., POM), thereby modifying $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ values relative to surface producers (Saino and Hattori, 1980; Pavlidou et al., 2020; Chaikalis et al., 2021; Pasqual et al., 2015). Such variability cannot be separated from surface signals under the available sampling design. Limited vertical resolution further constrains the ability to resolve these effects, which would likely require fully three-dimensional modeling approaches.

We argue that, taken together, these factors result in a low signal-to-noise ratio, in which spatially structured variation is obscured by unresolved local and non-spatial variability. The short spatial ranges and weak predictive performance observed across models therefore reflect the interaction between structured sources of variation and a sampling design that is not aligned with them, rather than the absence of isotopic gradients in the system.

Importantly, these limitations do not reflect poor data quality. The compiled dataset integrates high-quality observations from numerous studies (Table S1) and is large relative to many existing marine isoscapes (> 6000 samples from > 600 sites), comparable to or exceeding that of datasets used in isoscapes for the Southern Ocean ($N > 3000$ samples; St John Glew et al., 2021), North Atlantic ($N = 570$; Espinasse et al., 2022), or UK shelf seas ($N = 627$; St John Glew et al., 2019b). Metadata incompleteness or inconsistency should not be interpreted as a deficiency: in studies not designed for spatial analysis, such information is often unnecessary. Similar gaps are widespread – for example, Skinner et al. (Skinner et al., 2022) reported that 17% of isotope studies in coral reef ecosystems lacked sufficient detail to identify sampling time or location. Rather the limitation lies in the lack of alignment between sampling design and the requirements of spatial modeling. There is clear

structure in the data – and thus potential to generate informative isoscapes – but the opportunistic nature of sampling prevents meaningful separation of spatial and non-spatial components of isotopic variation.

More broadly, this case illustrates a general limitation of applying variance-partitioning spatial models to opportunistically compiled datasets. Isoscapes can be derived when sampling designs allow key sources of variation to be represented across space in a way that enables their effects to be distinguished statistically. When this condition is not met, variance is partitioned according to the structure of the data rather than underlying processes, and resulting isoscapes cannot be interpreted as true geographic signals. These challenges are not unique to the Mediterranean, but are likely to arise in any system where sampling is unstructured relative to spatially and ecologically structured sources of variation. For spatial isotope applications, reference datasets must be built from spatially explicit, coordinated, and appropriately replicated sampling efforts – such as basin-scale epipelagic surveys or Continuous Plankton Recorder-type programs (Richardson et al., 2006). High-quality data alone are not sufficient; alignment between sampling structure and modeling requirements is essential.

4.3. Implications for spatial inference from compiled datasets and future sampling design

Broad-scale gradients in $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ values have been successfully identified in several marine systems using opportunistic or mixed-source datasets (e.g., McMahon et al., 2013; Espinasse et al., 2020; St John Glew et al., 2021; Espinasse et al., 2022). However, these cases were generally more amenable to spatial modeling than the Mediterranean Sea. Samples were typically more homogenous and often restricted to specific epipelagic compartments (e.g., POM; St John Glew et al., 2021; copepods; Espinasse et al., 2020, 2022; or gelatinous plankton; St John Glew et al., 2019b), and in some cases were explicitly collected for isoscape construction (St John Glew et al., 2019b). This allowed models to more effectively separate spatial structure from species- or time-related effects and recover coherent isotope patterns. In contrast, the Mediterranean dataset integrates a broader range of sample types across an environmentally complex and dynamic basin. While this increases ecological relevance and spatial coverage, it also introduces multiple, interacting sources of variability that limit the ability of spatial models to accurately partition variance.

These findings have broader implications as compiled isotope datasets become increasingly available (e.g., Berto et al., 2025; IsoBank, 2025; IsoArch, 2025; see also Arnoldi et al., 2023; Silva et al., 2022) and are applied to spatial modeling (Perga et al., 2026). When key sources of variance are uncontrolled or spatially confounded, even advanced Bayesian spatial models cannot compensate for limitations in sampling design. This has consequences not only for isoscape construction but for any analysis relying on hierarchical or spatial inference from opportunistic data.

Several diagnostic indicators can help assess whether a dataset is suitable for spatial applications: (i) semivariograms showing only short-range autocorrelation; (ii) spatial fields with short estimated ranges and high variances; (iii) random effects (e.g., category) with posterior distributions that deviate from ecological expectations; and (iv) environmental covariates that fail to improve predictive performance or to reduce the spatial field's range and variance. Together, these signals indicate that spatial isotope structure is unresolved. In such cases, isoscape predictions should not be extrapolated beyond sampled locations, and even local predictions should be interpreted cautiously, as the spatial field may primarily reflect non-attributable variance rather than genuine local signals. More broadly, these limitations extend beyond spatial analyses. In non-spatial analyses (e.g., mixing models, niche analyses, or time-series), it may be even more difficult to determine whether a lack of predictable structure reflects true ecological variability or an artefact from uncontrolled variance in opportunistic datasets.

To support robust spatial inference, sampling design must be aligned with the structure of isotopic variation. This does not require uniform coverage, but key sources of variability must be represented across space in a way that allows their effects to be distinguished statistically. In practice, this means ensuring that different levels of important factors (e. g., category, season, depth, etc.) are not systematically clustered in specific regions or confounded with spatial gradients.

Future efforts in the Mediterranean and similar systems should prioritize systematic, targeted, and coordinated sampling strategies. To separate spatial from non-spatial effects, key sources of variance should be represented across space at scales comparable to observed or expected autocorrelation ranges. Replication does not need to occur at every location, but variation must be structured such that contrasts can be estimated independently of location. Stratified designs (e.g., inshore vs offshore systems; see [Fanelli et al., 2022](#); [Tesán-Onrubia et al., 2023](#)) may also be effective when environmental drivers operate at different scales. Without such structure, increasing sample size alone will not improve inference, as unresolved variability will continue to be absorbed into spatial components. In contrast, fine-scale, structured sampling – particularly in spatially heterogeneous areas such as coastal and island systems – would allow non-spatial variance to be explicitly modeled, reduce overfitting by the spatial field, and improve the detection of meaningful gradients. That said, resolving spatial variation across multiple scales simultaneously remains logistically constrained: capturing kilometer-scale variability across the entire Mediterranean would require sampling densities that are not realistically achievable. Where full standardization is not feasible, comprehensive, high-quality metadata remain essential to ensure that data structure can be incorporated into models.

Large-scale coordinated programs such as [GEOTRACES \(2025\)](#) demonstrate the feasibility of spatially extensive, intercalibrated sampling. In the Mediterranean, similar progress could be achieved by integrating isotope analyses into existing monitoring programs (e.g., MEDITS; [Spedicato et al., 2019](#); MEDIAS; [Giannoulaki et al., 2021](#); PELMED; [Ifremer, 2022](#)). Embedding isotope sampling within ongoing surveys would also align with ecosystem-based monitoring frameworks ([de Boois, 2019](#)). Adding $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ analyses to routine sampling of low-mid trophic level organisms – such as benthic filter feeders (e.g., bivalves; [Jennings and Warr, 2003](#); [Liénart et al., 2025](#); [Barnes et al., 2009](#); and pelagic species, e.g., forage fish, scyphomedusae; [St John Glew et al., 2019b](#)) – would provide spatially coherent proxies for baseline variation at relatively low cost, complementing efforts under Descriptor 4 (“Marine Food Webs”) of the EC Marine Strategy Framework Directive (MSFD; [EU, 2008](#)) to assess trophic structure. Such taxa offer advantages for isoscape development: primary producers and POM are often too dynamic to resolve large-scale patterns, zooplankton often include mixed trophic strategies, and higher trophic levels integrate signals over too broad spatial and temporal scales. In contrast, shorter-lived forage fish and similar organisms provide a balance between spatial coverage, ecological integration, and limited mobility ([Rumolo et al., 2025](#); [Ortiz et al., 2024](#)).

Ultimately, progress toward reliable isoscapes in the Mediterranean region and elsewhere will depend not on data volume alone, but alignment between sampling strategies and modeling objectives.

5. Conclusions

We compiled thousands of stable isotope measurements from across the Mediterranean Sea, revealing substantial isotopic variation and spatial structuring at multiple ecological and spatial scales across the region. These patterns indicate that meaningful isotopic heterogeneity exists within Mediterranean food web baselines and confirm that isotopic variation in the basin contains a spatially informative signal. However, advanced Bayesian models were unable to reliably resolve this heterogeneity into a single continuous isoscape from the available opportunistic dataset, despite its size and the high quality of the

individual studies. This does not imply isotopic homogeneity or the absence of spatial gradients across the Mediterranean, but rather an inability to resolve and reliably generalize those gradients given the mismatch between sampling design and the requirements of spatial inference. Our results demonstrate that even sophisticated hierarchical models do not yield reliable isoscapes when sampling is not structured relative to the underlying spatial and ecological gradients ([Rumolo et al., 2025](#)).

The compiled dataset nevertheless represents a valuable resource and diagnostic framework. Beyond documenting basin-wide isotopic variability, it provides empirical insight into the conditions under which spatial inference succeeds or fails, and therefore informs future sampling strategies. Assembled from thousands of measurements spanning the Mediterranean basin, it represents one of the most comprehensive stable isotope syntheses currently available for the region. Reliable isoscape development will require systematic, spatially explicit, and targeted sampling designs integrated with coordinated monitoring programs to constrain key sources of variability. In the Mediterranean, this challenge is amplified by oceanographic and ecological features that promote strong fine-scale variability, including oligotrophy, restricted circulation, localized inputs, mesoscale processes such as fronts, eddies, and upwelling, and complex habitat mosaics ([Darnaude et al., 2004](#); [Fanelli et al., 2011](#); [Rumolo et al., 2018](#); [Cartes et al., 2024](#); [Bury et al., 2024](#)), representing an ecological limit to the generalization of isotopic baselines. More broadly, as large datasets become increasingly available, caution is warranted (see [Kerr et al., 2025](#)). Opportunistic datasets are often biased ([St John Glew et al., 2019b, 2021](#); [Espinasse et al., 2022](#); [Courtillot and Rousset, 2017](#)), and sophisticated hierarchical models are no substitute for well-structured sampling. Robust ecological inference therefore depends not only on data availability, but on critically evaluating dataset structure and aligning sampling design with analytical objectives.

6. Data availability

All data and R code used in this study are publicly available on Zenodo ([Magozzi and Cobain, 2026](#)).

CRediT authorship contribution statement

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Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supplementary data

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